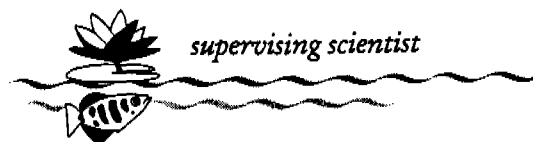




**Preliminary assessment
of petroleum
hydrocarbons in water
and sediment at Yellow
Water, Kakadu National
Park**

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C Camilleri
C Turley

April 1998



**Preliminary assessment of petroleum hydrocarbons in water
and sediment at Yellow Water, Kakadu National Park**

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Summary

In 1996, the environmental impacts of a commercial tour boat operation at Yellow Water, a highly valued wetland ecosystem within the World Heritage listed Kakadu National Park were assessed. While not specifically addressing pollutant-related issues associated with boating, the report did recognise the potential for pollution from outboard motors to impact upon the Yellow Water environment.

Leaking fuel and oil from outboard motors results in the contamination of waterways with petroleum hydrocarbons. The period of high boating activity at Yellow Water coincides with that of low water levels and no freshwater input, suggesting that petroleum hydrocarbons could accumulate to potentially toxic concentrations. Therefore, the present study aimed to provide a preliminary assessment of outboard motor-derived petroleum hydrocarbon levels in water and sediment during the late-dry season (ie a worst-case-scenario). In the event of petroleum hydrocarbons being detected, an assessment was to be made regarding the potential impacts on the Yellow Water environment, and the requirement for further monitoring/toxicity/risk assessment programs.

Water and sediment samples were collected from three sites at Yellow Water on December 16, 1997. The sampling sites were chosen according to where boating activity was considered to be greatest. Samples were analysed for benzene, toluene, ethyl-benzene and xylene (BTEX), and total petroleum hydrocarbons (TPH).

Petroleum hydrocarbons were detected in water samples from one of the three sites, at a concentration of 200 µg/L. Analysis revealed that the hydrocarbons were relatively high molecular weight compounds, possessing carbon numbers of between 15 and 28. Petroleum hydrocarbons were not detected above the Practical Quantitation Limit in any of the sediment samples.

For the contaminated site, the fact that no low molecular weight, highly volatile hydrocarbons were detected indicated that i) the contaminant event was not particularly recent, and ii) the source was oil, not gasoline (ie unleaded petrol).

Based on previous literature on the aquatic toxicity of petroleum hydrocarbons, it appears unlikely that acute toxicity would be observed for the plant and animal species inhabiting Yellow Water. However, at the concentration detected, there appears to be a possibility of sub-lethal effects to aquatic organisms.

The present study did not represent the 'worst-case-scenario' situation initially anticipated. Boating activity in December is much lower compared to the July – September peak, while Yellow Water had already received significant freshwater input prior to sample collection, potentially diluting hydrocarbon concentrations in the water.

As a result of the potential for sub-lethal toxicity to aquatic organisms, and also for the presence of higher hydrocarbon concentrations during peak boating periods, the implementation of a more comprehensive monitoring program, incorporating sampling periods before, during, and after the peak boating activity, is recommended. Such a program will better characterise the input of petroleum hydrocarbons to Yellow Water throughout the dry season. In addition, more specific chemical analyses, and potentially, toxicity assessments, will assist in the prediction of adverse environmental effects associated with boating activities at Yellow Water.

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1 Introduction

Yellow Water is a biologically diverse freshwater wetland situated within Kakadu National Park. Along with all wetland habitats in Kakadu, it is listed under the Convention on Wetlands of International Importance Especially as Waterfowl Habitat. Yellow Water possesses many environmental and cultural values, and as a result, has become a major tourist destination in northern Australia. The wetlands of, and surrounding, Yellow Water are a highly important habitat for waterfowl, with over 60 species, including up to a million magpie geese (*Anseranus semipalmata*) inhabiting the region at various stages during the annual seasonal cycle (Morton & Brennan 1991, Whitehead & Chatto 1996). A large tour boat operation allows visitors to view the abundant wildlife, including the birdlife and estuarine crocodiles, while also learning about the region's cultural history (Braithwaite et al 1996). In addition to its appeal as a sightseeing destination, Yellow Water is also a popular destination for recreational fishermen, in search of the sport fish, barramundi (*Lates calcarifer*).

In 1995, an extensive environmental, social and economic assessment of tour boat operations at Yellow Water was undertaken (Braithwaite et al 1996). While the study assessed environmental factors such as animal behaviour in response to tour boat operations, it did not assess the environmental effects of pollutant impacts arising from boating operations (ie pollution from outboard motors). However, Braithwaite et al (1996) did recognise that pollution, in the form of leaking fuel and oil from outboard motors, particularly two-stroke engines, was visible, and represented a potential environmental concern.

Fuel (gasoline) and oil (petroleum) contain a large number of different types of hydrocarbons, known collectively as petroleum hydrocarbons. Petroleum hydrocarbons can enter aquatic systems either as components of uncombusted petroleum and petroleum products (eg fuel, lubricating oils), or as products of incomplete combustion in exhaust emissions (Müller 1987). They have been shown to be quite toxic to a variety of aquatic fauna and flora (see reviews by Müller 1987 and Shales et al 1989), while they have also been found to accumulate in fresh and marine waters and sediments at elevated concentrations as a result of boating activities (Montz et al 1982, Mastran et al 1994, Trapido & Veldre 1996). The majority of the hydrocarbons in gasoline (eg unleaded petrol) are of low molecular weight and highly volatile (Müller 1987, Budavari 1996). While being quite toxic, such compounds evaporate from water quite rapidly (Robotham & Gill 1989, Mastran et al 1994), and hence do not usually represent great concern to the aquatic environment. In contrast, many hydrocarbons found in oil are of higher molecular weight and less volatile, and thus tend to persist in the water or deposit in the sediments (Mastran et al 1994). Two-stroke engines, used by most recreational boaters at Yellow Water are of particular concern as they contain considerable quantities of oil. It has been estimated that up to 56% of fuel used in two-stroke engines may be discharged directly into the water column, although an average of 10–20% is considered more realistic (Muratori 1968, Jackivicz & Kuzminski 1973). In addition, it has been estimated that the total discharge of hydrocarbons from an outboard two-stroke motor operating for one day would be equivalent to the sewage produced by a population of 400 people (Jackivicz & Kuzminski 1973). The above points emphasised the need for initial concern over the levels of petroleum hydrocarbons in both water and sediment at Yellow Water.

Yellow Water experiences the same seasonal hydrological conditions as most other waterbodies in the wet-dry tropics of Australia (see Taylor & Tulloch 1985, Finlayson et al

1989). Large amounts of rainfall during December to March (~1500–2000 mm) result in extremely large volumes of water at high flow rates passing into and through Yellow Water, as it connects with the main body of the nearby South Alligator River. During the Dry season, from May to approximately October, the water level declines, and flow ceases. If hydrocarbon levels do accumulate in the water and sediment at Yellow Water, it is likely that their concentrations would peak in the late Dry season, with the water at its lowest, the area having received little or no freshwater input for several months, and boating pressure having been maintained throughout the dry season. Samples taken during this period would most likely represent a 'worst-case-scenario' assessment of hydrocarbon levels at Yellow Water, and would assist in determining whether more extensive monitoring/toxicity/risk assessment programs were warranted.

Aims

The aims of the study were:

- i) to determine whether petroleum hydrocarbons could be detected in water and sediment at Yellow Water during the late dry season, as a result of both recreational and tourism boating operations,
- and if so;
- ii) to assess the requirements for further research on the potential ecological risks of petroleum hydrocarbons at Yellow Water.

2 Materials and Methods

2.1 Sampling sites

Approval for carrying out research at Yellow Water was granted by Parks Australia North and traditional owners on 12 December 1997 (a copy of the research permit is shown in Appendix A).

Water and sediment samples were collected between 1330 and 1530 on 16 December 1997, from three sites at Yellow Water. As financial constraints limited the size of the sampling program, the sites were selected according to where boating activity/pressure was considered to be greatest. The three sampling sites, all of which were located within Yellow Water billabong, are shown in figure 1. The type and number of samples collected from each site are outlined below, in table 1.

Table 1 Sampling sites, and type and number of samples to be collected from Yellow Water

Sampling site	Number of samples		
	surface water	sediment	Total
<i>A: public boat launching area</i>	2	2	4
<i>B: tour boat embarkation area</i>	2	2	4
<i>C: Yellow Water billabong - South Alligator River junction</i>	2	2	4
			12

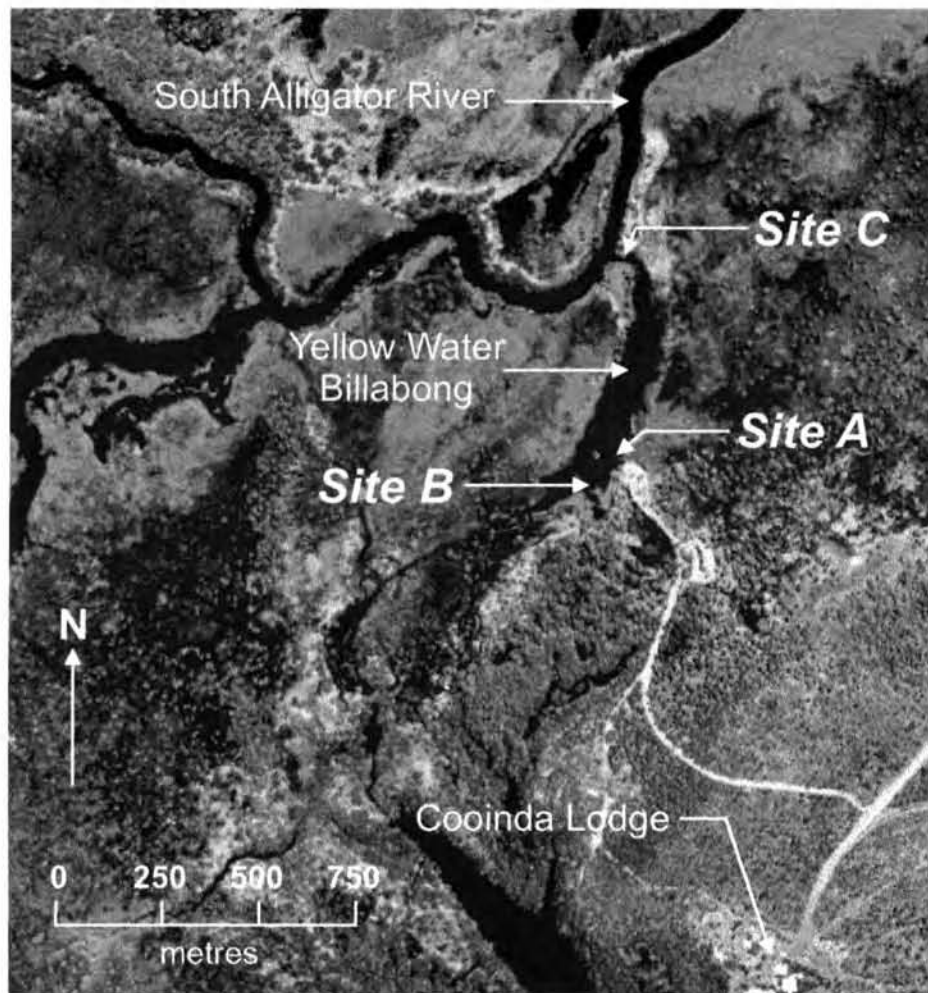


Figure 1 Aerial photograph of Yellow Water during the dry season, showing the three sampling sites. Site A: public boat launching area; Site B: tour boat embarkation area; Site C: Yellow Water billabong - South Alligator River junction (Photograph supplied by G. Lindner, Parks Australia North)

2.2 Sample collection

Samples were collected from a boat powered by a four-stroke outboard motor. Care was taken to switch off the motor well before reaching the sampling site to avoid the chance of contamination from the motor. Water was collected in 500 ml Schott bottles that had been washed in 5% HNO₃ and dried the day before sampling. Two surface water samples were taken at each site. Sediment was collected using a long PVC rod with an aluminium scoop fastened to one end. Sediment was placed in 250 ml Schott bottles prepared as described above. Two sediment samples were collected at each site. A clean water blank (Milli-Q water) and sediment blank (empty bottle) were also included for analysis.

For all samples, care was taken to minimise the air space in the neck of the bottle, so as to avoid evaporation of the more volatile hydrocarbons. Gloves were worn during all sample collections. Aluminium foil liners were placed over the openings of the bottles and the screw-cap tightly fastened. Due to the unavailability of teflon-lined screw caps, the water samples were not acidified upon collection. Following labelling, the samples were placed on ice in the dark until transportation to the laboratory, where they were placed in the dark at 4°C, overnight. A separate water sample was also taken from within Yellow Water Billabong for measurement of pH and conductivity.

2.3 Sample analysis

On 17 December the samples were sent by air to a NATA-registered analytical chemistry laboratory in Sydney (AMDEL) for analysis. Due to flight delays, the samples arrived at the laboratory on the morning of 19 December, approximately 24 h later than expected. All samples were analysed for benzene, toluene, ethylbenzene and xylene (BTEX), and total petroleum hydrocarbons (TPH), using gas chromatography (AMDEL Methods: E084, Total Petroleum Hydrocarbons in Soil; E083, Total Petroleum Hydrocarbons in Water; E054, Benzene, Toluene, Ethylbenzene & Xylene in Soil; E052, Benzene, Toluene, Ethylbenzene & Xylene in Water).

3 Results

3.1 Water samples

The QA/QC report for the analyses is presented in Appendix B. The pH and conductivity of Yellow Water billabong water at the time of sampling were 7.6 and 50.5 µS/cm, respectively.

Petroleum hydrocarbons (BTEX or TPH) were not detected above the Practical Quantitation Limits (PQL)¹ in water samples collected from the boat launching area (Site A), the tour boat embarkation area (Site B) (table 2), or in the clean water blank. However, petroleum hydrocarbons were detected in both water samples taken from the Yellow Water billabong-South Alligator River junction (Site C) (table 2). Analysis revealed the hydrocarbons possessed carbon numbers of between 15–28 (table 2). The analysis report for all the water samples is presented in Appendix C.

¹ The PQL is typically 2-10× the method detection limit.

Table 2 Levels of BTEX and TPH in water at three sites at Yellow Water, Kakadu National Park. Measurements are in µg/L (parts per billion; ppb)

	Site*					
	A		B		C	
	Rep. 1	Rep. 2	Rep. 1	Rep. 2	Rep. 1	Rep. 2
Benzene	nd	nd	nd	nd	nd	nd
Toluene	nd	nd	nd	nd	nd	nd
Ethylbenzene	nd	nd	nd	nd	nd	nd
Xylene	nd	nd	nd	nd	nd	nd
TPH:						
C₆-C₉	nd	nd	nd	nd	nd	nd
C₁₀-C₁₄	nd	nd	nd	nd	nd	nd
C₁₅-C₂₈	nd	nd	nd	nd	200	200
C₂₉-C₃₆	nd	nd	nd	nd	nd	nd
Total (C₆-C₃₆)	-	-	-	-	200	200

* Site A: private boat launching area; Site B: tour boat embarkation area; Site C: Yellow Water billabong - South Alligator River junction

nd: not detected

3.2 Sediment samples

No petroleum hydrocarbons (BTEX or TPH) were detected above the PQL in any of the sediment samples, including the sediment blank. The analysis report for the sediment samples is presented in Appendix C.

4 Discussion

4.1 Source of petroleum hydrocarbons at Yellow Water

As stated in section 2.1, the Yellow Water sampling sites were chosen because they represented the areas of greatest boating activity. Of the three sites, the private boat launching area and the tour boat embarkation area (sites A and B, respectively) were considered most likely to be contaminated. However, it was the water samples from site C, the Yellow Water billabong – South Alligator River junction that revealed the presence of petroleum hydrocarbons, at a concentration of 200 µg/L. The junction is a narrow entrance into the South Alligator River, through which the majority of vessels travelling throughout Yellow Water pass. The chemical analyses provided some information on the potential source of the hydrocarbons. The fact that no lower molecular weight, highly volatile hydrocarbons (eg BTEX) were detected in the sample suggested that the contamination event was not particularly recent. This effectively eliminated the contaminant source being the outboard motor of the sampling boat. In addition, that only large carbon chain hydrocarbons were detected (ie C₁₅ – C₂₈), indicated that the source was oil, and not gasoline (ie unleaded petrol), which characteristically contains a mixture of C₄ to C₁₂ hydrocarbons (Müller 1987, Budavari 1996).

The fact that higher molecular weight hydrocarbons were detected in surface water but not in sediment was somewhat unusual, as such compounds tend to deposit in sediments as a result of their hydrophobicity and high octanol-water partition coefficient ($\text{Log}K_{ow}$) (Mastran et al 1994). Illustrating this property, Mastran et al (1994) reported total concentrations of polycyclic aromatic hydrocarbons (PAHs) in surface water and sediment of the Occoquan reservoir in Virginia of up to 18 $\mu\text{g/L}$ (ppb) and 1319 $\mu\text{g/kg}$ (ppb), respectively, a difference of approximately two orders of magnitude. However, most pollutants also have a tendency to concentrate in the surface microlayer (ie top 200 μm) of a water body (Mullins 1977), thus potentially explaining the presence of hydrocarbons in surface water but not sediment. While multiple depth sampling would have resolved this, the budget of the sampling program did not allow for it. It should also be noted that increased water temperatures, high dissolved organic matter (DOM) content, and low salinity, all common characteristics at Yellow Water, serve to increase the solubility of hydrocarbons (Müller 1987). Alternatively, the hydrocarbons may not have been dissolved in the water column, but adsorbed to suspended sediment/particulates within the water column. Supporting this, a recent study reported that total PAHs in marine water were mostly bound to particulates than dissolved in the water column (Zeng & Vista 1997). Filtration of Yellow Water water samples prior to analysis would reveal whether the above hypothesis was correct.

4.2 Aquatic toxicity of petroleum hydrocarbons

The toxicity of petroleum products to aquatic organisms varies enormously. For example, Tsvetnenko (in press) compiled toxicity data for marine species on different petroleum products including crude oils, diesel oils and bunker fuels, with LC_{50} values (based on μg petroleum hydrocarbons/L) ranging from 190 $\mu\text{g/L}$ for the ghost crab, *Ocypode quadrata*, to 2620 $\mu\text{g/L}$ for the alga, *Isochrysis* sp. The advisable water quality criterion (AWQC) derived from the toxicity values (based on US EPA Guidelines) was 7 $\mu\text{g/L}$ (Tsvetnenko in press), well below the level of petroleum hydrocarbons found at Yellow Water. However, this value was derived for marine waters only, and thus does not necessarily apply to freshwater ecosystems. Müller (1987) and Shales et al (1989) reviewed the toxicity of oils and petroleum products to freshwater organisms. However, much of the data are difficult to compare to the levels detected at Yellow Water because they involved crude oils with complex, and often uncharacterised compositions. In addition, the hydrocarbon content, and potentially toxicity, of petroleum products such as two-stroke outboard motor fuel does not necessarily reflect that of the parent crude oil (Gill & Robotham 1989). Nevertheless, a summary of the aquatic toxicity of petroleum and petroleum products is presented below.

Apparently, microorganisms in freshwater environments exhibit no wide-scale toxicity to petroleum hydrocarbons, with many responses actually involving increases in microbial populations (Shales et al 1989). Kauss et al (1973, as cited by Shales et al 1989) reported a 40% decrease in algal cell number compared to controls for the green alga, *Chlorella vulgaris*, two days following exposure to a 90% extract of outboard motor oil, indicating some toxic effect. However, cell number had returned to normal after four days' exposure. King & Coley (1985) exposed three species of duckweeds (*Lemna* spp.) to water soluble fractions (WSF) of various oils. Toxicity varied enormously depending on the oil type and the species of *Lemna*, from no effect at 100% WSF of oil, to 100% growth inhibition at 10% WSF of oil. The 48 h LC_{50} of the WSF of a crude oil to the freshwater invertebrate, *Asellus aquaticus*, was reported to be 11.58 mg/L (Ramusino & Zanzottera 1986). However, the ephemeropterans, *Baetis rhodani* and *Ecdyonurus helyeticus*, were more sensitive, with 100% mortality occurring at 10 mg/L WSF (Ramusino & Zanzottera 1986). Woodward et al

(1987) reported up to 90% mortality for the invertebrate genera, *Baetis* and *Isoperla*, at 0.5–0.7 mg/L WSF of several crude shale oils. In addition, the 96 h LC50s of the same WSFs to the cut-throat salmon, *Salmo clarki*, ranged from 1.3–2.1 mg/L (Woodward et al 1987). Müller (1987) summarised the literature on low level, sub-lethal toxicity of hydrocarbons to freshwater species. Effects of various individual hydrocarbons on fish and crustaceans were reported at concentrations of 0.1–20 µg/L and 10 - 80 µg/L, respectively. In addition, toxic effects of a crude oil WSF to the brown alga, *Fucus edentatus*, have been reported at 0.2 µg/L (as reviewed by Müller 1987).

In addition to aquatic animals, birds may also be at risk from waters contaminated with petroleum hydrocarbons. In the case of chronic hydrocarbon pollution, such as that arising from boating activity (excluding large spills), birds could be exposed directly, by drinking contaminated water (Shales et al 1987), or via biomagnification, having eaten contaminated food (ie fish, invertebrates, plant matter) (Roper et al 1996). Brunström (1991) investigated the toxicity of various PAHs to avian embryos. A 200 µg/kg (ppb) dose of benzo[k]fluoranthene, a PAH known to be present in many crude oils (Müller 1987), resulted in 100, 65 and 44% mortality of domestic duck (*Anas platyrhynchos*), turkey (*Meleagris gallopavo*), and eider (*Somateria mollissima*) embryos, respectively (Brunström 1991). A 2000 µg/kg mixture of 18 PAHs was also highly toxic, resulting in 100, 83 and 94% mortality of embryos of the above three species, respectively (Brunström 1991). These results confirm the susceptibility of avian early life stages to petroleum hydrocarbons.

When considering the toxicity of petroleum hydrocarbons, two additional factors deserve recognition: i) the production of highly toxic metabolites of hydrocarbons, and ii) the high carcinogenicity of certain PAHs known to be present in oil and exhaust emissions. Highly toxic metabolites of PAHs are known to be produced as a result of their exposure to light (photooxidation; Müller 1987, McConkey et al 1997), while the metabolism of many PAHs by the mixed function oxidase enzyme system within living organisms results in the formation of highly carcinogenic metabolites (Ahokas 1991). Depending on the nature of the hydrocarbons present at Yellow Water, the above two factors could be of concern.

4.3 Implications for the Yellow Water environment

While it is difficult to predict potential impacts on the Yellow Water environment without further knowledge of the identity of the C15–C28 hydrocarbons detected in the present study, some general conclusions can be made. Based on the levels of petroleum hydrocarbons detected in the present study and the acute toxicity data summarised above, it appears unlikely that acute toxicity would be observed for local animal and plant species at Yellow Water. However, the potential for sub-lethal toxicity due to continual low-level hydrocarbon contamination of Yellow Water may represent a concern, although this may be somewhat mitigated by annual flushing of the system during the wet season.

The intention of the present study was to assess hydrocarbon levels during the late-dry season (ie early December), when it was thought hydrocarbon build-up would be at its greatest. However, tour boat activities at Yellow Water during November–December (average of 243 cruises per month) average only 50% of the July–August peak period (average of 512 cruises per month) (Braithwaite et al 1996). It is likely that private boat usage also reflects that of the tour boat operations, resulting in the vast majority of boating activity at Yellow Water occurring during the months of July–September. Although the volume of water within the system is far greater at this point than during the late-dry season, it is still highly possible that levels of petroleum hydrocarbons will exceed those detected in the present study. Supporting

this, periods of peak boating activity have previously been shown to be associated with elevated PAH concentrations (Mastran et al 1994, Trapido & Veldre 1996). In addition, the samples in the present study were taken approximately two weeks later than was considered ideal, with Yellow Water having already received significant volumes of inflowing water from Jim Jim Creek and the South Alligator River (P. Shaughnessy, Senior Ranger Jim Jim District, pers. comm.). While this was not likely to have affected sediment composition, petroleum hydrocarbons in the water column may have been substantially diluted.

The fact that petroleum hydrocarbons were detected at only one site indicated that contamination of Yellow Water was not widespread, instead, occurring in localised areas as a result of fuel leakage and exhaust emissions from outboard motors. In addition, exposure to high UV radiation levels and a diverse microbial population, factors that assist in the degradation of petroleum hydrocarbons in the aquatic environment (Robotham & Gill 1989), would decrease the risk of widespread contamination. However, it is still possible that contamination during the peak boating periods may be greater and more widespread, as has previously been demonstrated for other, larger freshwater bodies (Mastran et al 1994; Trapido & Veldre 1996).

5 Conclusions and Recommendations

Petroleum hydrocarbon contamination, albeit at low levels, was found to occur at Yellow Water, during a period which, in hindsight, did not necessarily represent the 'worst-case-scenario' initially anticipated. At the time of sampling, boating activity was relatively low, while Yellow Water had already received significant freshwater input, yet petroleum hydrocarbons were detected at levels that have previously been shown to result in sub-lethal toxicity to aquatic organisms. As a result, the development and implementation of a more comprehensive and statistically sound monitoring program, focussed on periods of peak boating activity is recommended. Periodic collection and analysis of water (surface water and water column) and sediment samples should be carried out prior to, during, and after the period of peak boating activity (July - September), in order to detect whether a significant boating-related increase in contamination exists, and if so, to monitor the persistence of contamination as boating activity decreases.

In addition, further chemical characterisation/identification of the petroleum hydrocarbons present in water or sediment at Yellow Water should be carried out, to assist in the prediction of the potential adverse effects on the aquatic environment. Finally, given significant and consistent levels of contamination, an assessment of the toxicity of relevant petroleum hydrocarbons, or contaminated Yellow Water water on local aquatic organisms, using standard laboratory toxicity testing procedures would also assist in predicting the risk of adverse environmental effects associated with boating activities at Yellow Water.

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Appendix A Copy of permit to carry out research at Yellow Water

RECEIVED

16 DEC 1997

58

**PARKS
AUSTRALIA
NORTH**

Kakadu National Park
Parks Australia North
PO Box 71
Jabiru NT 0886
Tel: 08 8938 1100
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Mr Rick van Dam
Environmental Research Institute
of the Supervising Scientist
Locked Bag 2
Jabiru NT 0886

Permit Number : RK 471

Date of Issue : 12 December 1997

**PERMISSION TO CONDUCT SCIENTIFIC RESEARCH
IN KAKADU NATIONAL PARK**

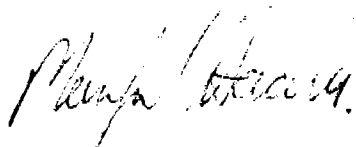
Permittee: Rick van Dam of the Environmental Research Institute of the Supervising Scientist is given permission under Regulation 27 of the National Parks and Wildlife Regulations to conduct scientific research in Kakadu National Park.

Research permitted: Sampling of water and sediment for hydrocarbons in Yellow Water

Period of permit: 15 to 19 December 1997

Access to specific locations requires the approval of relevant District Supervisor/s. Permit holders must obtain such approval before commencing work.

Subject to the conditions overleaf.



**DELEGATE
DIRECTOR OF NATIONAL PARKS & WILDLIFE**

FAKED
12.97

 **Environmen
Australia**
Biodiversity Group

**Appendix B Quality assurance/quality control report for
chemical analyses**

QA/QC APPENDIX NO. 7E02139

<u>Method</u>	<u>Description</u>
E052	Benzene, Toluene, Ethylbenzene & Xylene
E054	Benzene, Toluene, Ethylbenzene & Xylene
E083	Total Petroleum Hydrocarbons
E084	Total Petroleum Hydrocarbons

Chromatography QA/QC

	Yes	No	N/A
Retention Time Window Within Acceptance Criteria($\pm 2\%$)	✓		
Check Standard Within Acceptance Criteria($\pm 10\%$)	✓		
Recalibration Within Acceptance Criteria($\pm 15\%$)	✓		

Other QA/QC

Holding time conforming With Method Specification	✓
Chain of Custody Attached	✓

N/A = Not Applicable

Comments

1. Laboratory QA/QC including Duplicates, Matrix Spike Duplicates, and check/reference samples are included in this QA/QC appendix. (Where applicable)
2. Inter-Laboratory proficiency trial results available on request. (Where applicable)
3. Surrogate description and recoveries are recorded in the Report. (Where applicable)
4. Acceptance criteria for specific analytes are listed on each QA/QC page.
5. Practical Quantitation Limit (PQL is typically 2-10 x method detection limit (MDL)).
6. PQL's are matrix dependent and are increased accordingly where sample extracts are diluted.
7. Results are uncorrected for matrix spike or surrogate recoveries.



per G.W. ANDERSON
Manager Environmental Sydney

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Refer to Amdel-Sydney Quality Control Manual SPM-01 4th Edition 1/7/97

Appendix C Analysis report for all samples

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PQL = Practical Quantitation Limit

LNR = Samples Listed not Received

nd = Not Detected (<PQL)

-- = Not Applicable

Soils : mg/kg (ppm) dry weight unless otherwise specified

Waters : mg/L (ppm) unless otherwise specified

Leachates : mg/L (ppm) in leachate

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[illegible]

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Soils : mg/kg (ppm) dry weight unless otherwise specified

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Leachates : mg/L (ppm) in leachate

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plus Cover Page

plus Cover Page

[illegible]

Soils : mg/kg (ppm) dry weight unless otherwise specified

Waters : mg/L (ppm) unless otherwise specified

Leachates : mg/L (ppm) in leachate

* BTEX result taken from TPH, GC-FID run.

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[illegible]

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